

# Dust career impacts on *Pinus halepensis* growth

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**Abstract**— *Pinus halepensis* Mill., is a most common tree in the Mediterranean basin. In Tunisia, specifically Kroumirie, it is an excellence species. However, for several years, we assist a continual deterioration of this ecosystem type. Several factors are the origin for this degradation: insects and fungi attack, fire, aging populations, low regeneration and hardening climate. This degradation is further accentuated by installing careers around the pine forest. Our objective in this study was to identify the career dust influence on growth and productivity of Aleppo pine; through dendrochronological approach (tree rings study) and dendrometric approach (measurement of diameter, height and survival rate). Study is accomplished on two populations: a reference site 'Charchara' located away from mining and Oued el maaden site near a gravel extraction career. In addition, physical and chemical parameters are analysed on soil and water samples collected in the two sites. Results showed significant differences of parameters studied between stations both in dendrometric and dendrochronological parameters over time. Second, in physical and chemical parameters, a very high content on metals traces was found on soil and water in Oued el maaden site.

**Keywords**— career, growth, *Pinus halepensis*, productivity.

## Résumé - Impact des poussières de carrière sur la croissance de *Pinus halepensis*.

Le pin d'Alep '*Pinus halepensis* Mill' est une essence forestière la plus commune du bassin méditerranéen. En Tunisie, plus précisément en Kroumirie, elle constitue une essence forestière d'excellence. Toutefois, depuis plusieurs années, on assiste à une dégradation continue de ce type d'écosystème. Plusieurs facteurs sont à l'origine de cette dégradation : attaque d'insectes ou de champignons, incendies, vieillissement des peuplements, faible régénération et durcissement du climat. Cette dégradation serait encore plus accentuée par l'installation de carrières aux alentours de la pinède. Ce travail a donc pour objectif de déceler l'influence de poussière de carrière sur la croissance et la productivité du pin d'Alep ; par le biais de l'approche dendrochronologique (étude des cernes de croissance) et dendrométrie (mesure de diamètre, hauteur et taux de survie). Pour ce faire, deux sites sont choisis : un site de référence situé loin de la carrière 'Charchara' et un autre au voisinage de la carrière 'Oued el Maaden'. De même, des paramètres physicochimiques ont été mesurés dans des échantillons de sol et d'eau collectés au niveau des deux sites. Les résultats ont montré des variations significatives des paramètres étudiés entre les deux stations aussi bien en hauteur qu'en largeur et aussi au niveau de la croissance radiale au cours du temps. En deuxième lieu, au niveau des paramètres physicochimiques, une teneur très élevée en éléments traces métalliques a été identifiée sous pin d'Alep à Oued el maaden.

**Mots clés** : Carrière, croissance, *Pinus halepensis*, poussière, productivité.

## I. INTRODUCTION

Aleppo pine (*Pinus halepensis* Mill.) is a widely distributed species all around the Mediterranean Basin. It extends from Morocco in the west to Palestine and Jordan in the east, and from France in the north to Tunisia and Algeria in the south (Couhert *et al.*, 1993; Vila, 2008). It prefers regions with a high mean annual temperature and is adapted to prolonged summer droughts (Quezel, 1986). Given its plastic behavior towards climate and soil, the Aleppo pine is the most common tree species in Tunisia with a total area of 296 571 28 ha, representing over 35% of forest land in the country (DGF, 1995). The Kroumirie zone (the North West forest in Tunisia) is characterized by decreasing aridity of the climate from East-South-East to West-North-West. However, the Aleppo pine constitutes an excellence forest essence in this type of ecosystem.

Despite the conservation and protection efforts of this species, mainly used essentially for reforestation, we observe for several years a continual degradation of this essence. Among the degrading factors, the installation of careers near pine forests, accentuates the decrease of growth and the low natural regeneration of the species.

Human activities such as mining, industry, agriculture, waste treatment, and transportation release substantial amounts of trace elements into the environment (Nriagu and Pacyna, 1988; Nriagu, 1990a). Increasing anthropogenic influences on the

environment, especially dust careers, have caused negative changes in natural ecosystems; decreased biodiversity, simplified structure, and lowered productivity (Shparyk Y.S., and Parpan V.I., 2004). These degradation processes can be seen especially in forest ecosystems. Deterioration of forest health has been a major concern of the world community for the past 20-30 years (Royal Ministry for Foreign Affairs, 1971; UN/ECE, 1979; Smith, 1981). Dust carriers affect organisms within forest ecosystems, and inhibit decomposition of organic matter on forest floors, which disrupts nutrient cycles (Cotrufu *et al.*, 1995). Anthropogenic elements deposited to forest soils accumulate in the surface layers of the soil (Corwin *et al.*, 1999; Hou *et al.*, 2005a; Ruan *et al.*, 2008), and are usually immobilized there for a long time (e.g. Friedland and Johnson, 1985; Hawkins *et al.*, 1995).

In particular, smelting, incineration, and transportation release pollutants into the atmosphere. Transportation emits trace elements from fuel combustion (Huang *et al.*, 1994; Wang *et al.*, 2003). Thus, the air around career is rich in trace elements (Mizohata *et al.*, 2000). The fine particulate matter like An, As, Sb and Pb is suspended in air for several weeks and is often transported over large distances. The particulate matter is eventually deposited to soil either directly or via deposition onto vegetation (Takamatsu *et al.*, 2000; Sakata *et al.*, 2006). Therefore, pollution of soil and water by trace elements progresses gradually, even in montane forests far from urban areas.

Although there are no rules or guidelines on montane soil pollution in Tunisia especially in Kroumirie. It is important to analyze current levels of anthropogenic elements in montane forest and to determine potential effects of this pollution. However, such studies have rarely been carried out in Kroumirie and the large-scale impact of dust careers in forest ecosystems is not so evident. The aim of the studies presented here was to identify the impact of dust career emissions on the Aleppo pine forest. Two populations were targeted, the first is located in front of a career and the second has served as a control population.

## II. MATERIALS AND METHODS

### 2.1. Study area

Our study area was located in the Kroumirie zone (the north west of Tunisia) belonging to the Beja governorate characterized by a subhumid climate with mild winters and moderate hot summer. Two natural Aleppo pine populations are chosen: a reference site 'Charchara' (latitude 36°52'34,10''N, longitude 9°06'44,86''E, altitude 349m) located away from mining and Oued el maaden site (latitude 36°54'05,82''N, longitude 9°06'44,86''E, altitude 189m) near a career (Fig. 1). Sites are circular with an area of 500m<sup>2</sup>, that a circle of 12.5m of radius. To note that Oued el maaden population is in front of a career exploited for lead extraction Pb since the 1960 and actually for a gravel extraction since the 1980 (Fig. 2).



FIG. 1. MAP SHOWING SAMPLING LOCATIONS OF ALEPPO PINE POPULATIONS. ★ : CHARCHARA ; ● : OUED EL MAADEN.



**FIG. 2. OUED EL MAADEN CAREER.**

## **2.2. Data sampling and chronology construction**

### **Biometric parameters**

Biometric measurements were taken from all representative trees of each population (60 in Oued el maaden and 54 in Charchara). These parameters correspond to the total height of trees and the diameter growth at 1.30 m from the trunk base.

### **Soil and water sampling**

A soil profile was selected to be the most representative of each station. A large pit has been dug. Then, each soil horizon was refreshed by removing the soil which was in contact with the tools. A maximum of three representative horizons were sampled down the profile. Thereafter, water samples were collected from the existing sources in each site.

Only the fine fraction ( $< 2\text{mm}$ ) of representative samples from soil was chosen to be analyzed. To ensure the representativeness of the sample fraction to be analyzed, each soil sample was carefully separated into four equivalent subsamples. Then, two opposite subsamples were mixed and separated again into four subsamples. This procedure was repeated three times. Finally, the final representative subsample was carefully crushed in an agate mortar.

The main soil parameters (granulometry, pH, organic matter content OM, phosphorus extracted  $\text{P2O5}$ , content of metal traces ETM) have been analyzed at the INRGREF laboratory of soil Sciences following classical methods (Ponette et al., 1997).

For water samples, they were taken from the main sources at each site (watercourse in Charchara, river and watercourse in Oued el maaden). The analyses conducted aim to characterize the pH, electric conductivity e.c., suspended matter content MES and metal traces ETM parameters.

### **Dendrochronological approach**

Tree-ring width has long been used to reconstruct historical trends in tree vigor and has been shown to be responsive to various environmental factors (Fritts, 1976; Biondi, 1993). Tree rings, if validated as an environmental archive for pollution, would provide a convenient, geographically widespread archive for studying the temporal and spatial distribution of atmospheric pollutants. More recently, metal element concentrations in tree rings have been used as biomonitors of chemical parameters in the environment, including soil solution, chemistry and atmospheric pollution (Padilla and Anderson, 2002; Poszwa et al., 2004).

In our study, dendrochronology is used to determine the impact of dust carrier in growth of Aleppo pine population. We selected six individuals per population and collected cores with a Pressler borer at two levels of each tree (at 0.30m and at 1.30 m from the trunk base). The selected trees are settled in homogeneous stands, with dominant status, healthy, without significant deformation and are not subject to specific microclimatic conditions. After drying and polishing cores, in order to ensure the absence of anatomic abnormalities and to assign for each identifies the exact vintage of its formation, the tree-ring

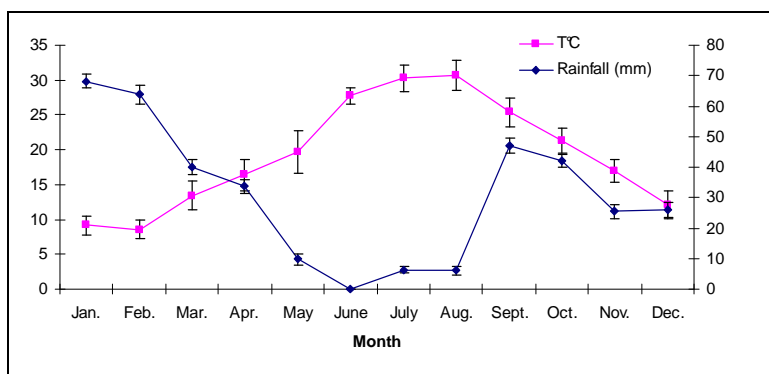
series were interdated (Fritts, 1976). For each population, the crossdating was achieved by seeking to identify the same ring sequences between the two cores of the same tree and between trees from the same plot. The cross dating being made, the thickness of the rings was measured with an accuracy of 1/1000 mm using the measuring table (LINTAB 5; Rinntech, Heidelberg, Germany), a stereoscope (MZ 6; Leica, Wetzlar, Germany) and the TSAP Win Scientific 4.63 soft-ware (Rinntech, 2009). The crossdating was checked visually by comparing the curves of the interannual variation of the rings thickness (Munaut, 1978). Then, to reduce longterm variations related to biological factors (geometric effect, age of individuals, changes in habitat), the sets of raw data were standardized which reduces the fluctuations of low and medium frequencies and lead to compare representative chronologies of each individual (Fritts, 1976). For this, a low-pass filter (window: 20 years) was used.

**2.3. Statistical analyses**

The relationships between dendrometric parameters (DBH, H) were described using the software STATITCF (ver.F). The measures were the object of an analysis of the variance to one factor following the case, significance levels were established at  $P < 0.05$ . It was completed by a multiple comparison of the averages by the test of Newman-Keuls test (at 5%) according Dagnelie (1986). The graphical exits are realized with the software Exel 2000.

**III. RESULTS**

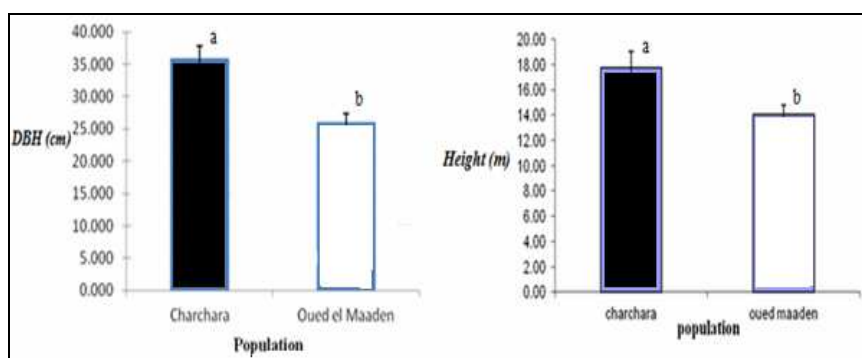
A regional climatic analysis was established for 2014. Data taken from button pile located near the two populations studied (Fig. 3).



**FIG. 3. AVERAGE MONTHLY TEMPERATURE AND RAINFULL AT THE BEJA GOVERNORATE.**

**Overall growth parameters**

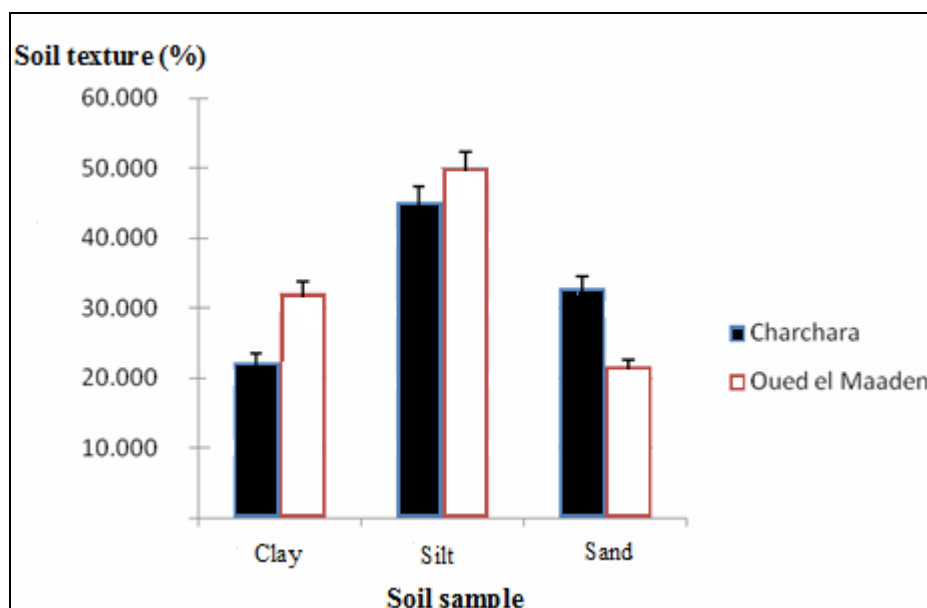
Overall growth differed significantly between the two sites (Fig. 4). On average, the trees in Charchara were about  $17.75 \pm 1.26$  m and  $39.83 \pm 3.00$  cm thicker in DBH. However, those of Oued el maaden were respectively  $14.08 \pm 1.00$  m and  $26.41 \pm 0.8$  cm.



**FIG. 4. OVERALL GROWTH PARAMETERS OF THE TWO POPULATIONS; GIVEN ARE MEAN VALUES ± STANDARD DEVIATION FOR EACH POPULATION. RESULTS OF THE NEWMAN-KEULS TEST CALCULATED FOR THE TWO SITES ARE INDICATED BY LETTERS.**

### Soil and water analyses

The texture of the soils is presented in Fig. 5. Results reveal highly significant values in clay ( $32.17\% \pm 5.29$ ) and silt ( $49.86\% \pm 3.59$ ) in the Oued el Maaden station. These proportions decrease in Charchara station ( $22.33\% \pm 0.83$  of clay and  $45.17\% \pm 2.3$  of silt).



**FIG. 5. SOIL TEXTURE OF THE TWO POPULATIONS. RESULTS INDICATE THE AVERAGE VALUE OF THE THREE HORIZONS FOR EACH SOIL SAMPLE (FINE FRACTION < 2 MM).**

The soil samples showed pH values which oscillate between 7.05 and 7.85 in different horizons to both profiles of the two populations (Table. 1). The average content of organic matter O.M and phosphorus extracted P<sub>2</sub>O<sub>5</sub> in Charchara soil samples were respectively  $4.5 (\%) \pm 1.5$  and  $116.87 (\text{ppm}) \pm 24.18$ . In Oued el maaden, the O.M content was  $4.4 (\%) \pm 0.3$  and the P<sub>2</sub>O<sub>5</sub> was  $44.21 (\text{ppm}) \pm 15.23$ .

**TABLE 1  
THE MAIN PARAMETERS ANALYSED OF THE SOIL SAMPLES IN THE TWO POPULATIONS.**

Populations	pH	Organic matter content O.M (%)	P <sub>2</sub> O <sub>5</sub>
Charchara	7.05	$4.5 \pm 1.5$	$116.87 \pm 24.18$
Oued el maaden	7.87	$4.4 \pm 0.3$	$44.21 \pm 15.23$

Table 2 shows the soil content of metal traces ETM in the two population profiles studied. In all horizons, the abundance order of metal trace content in soil samples is: Pb > Zn > Cd (Table. 2). This abundance was the same on the two sites. The highest values were 256.15 for Pb, 7.95 for Cd at depth-1 (0-10 cm) and 133.15 at depth-2 (10-30 cm) in Oued em maaden. There were highly significant average contents of Pb, Zn and Cd in Oued el maaden site ( $195.48 \pm 3.20$ ,  $87.24 \pm 6.74$ ,  $4.61 \pm 3.48$ ) compared to Charchara site ( $81.67 \pm 3.44$ ,  $37.15 \pm 5.24$ ,  $2.96 \pm 0.33$ ).

**TABLE 2**  
**CONTENT AND DISTRIBUTION OF METAL TRACE ELEMENTS IN THE SELECTED SAMPLES (FRACTION < 2 MM).**

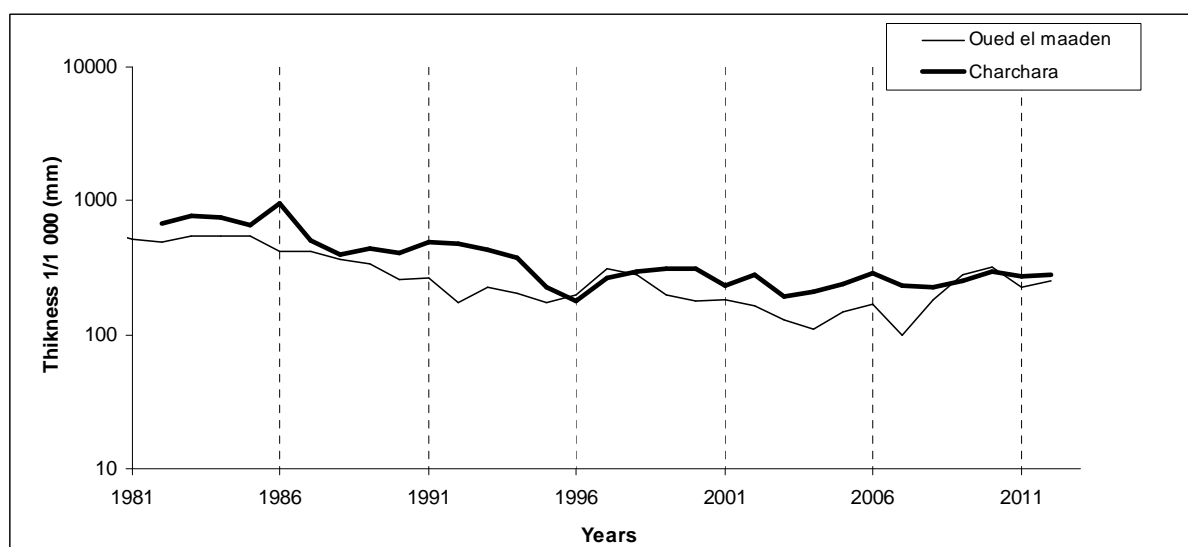
Station	Depth (cm)	Trace elements		
		Pb	Zn	Cd
Charchara	0 - 10	29.9	25.45	3.7
	10 - 25	218.2	52.6	1.09
	25 - 50	39.2	37.15	3.25
	50 - 80	39.4	33.4	3.8
Oued el maaden	0 - 10	<b>256.15</b>	49.47	<b>7.95</b>
	10 - 30	248.2	<b>133.05</b>	3.15
	30 - 70	82.1	79.21	2.75

For water analyses, there were no significant differences for pH values and electric conductivity 'e.c' between the two sites (Table .3). However, about the pollution level, the suspended matters content 'MES' and the metal traces parameters 'ETM' were highly significant in Oued el maaden compared to Charchara (Table .3).

**TABLE 3**  
**THE RELATING PARAMETERS TO THE WATER QUALITY IN THE TWO SITES.**

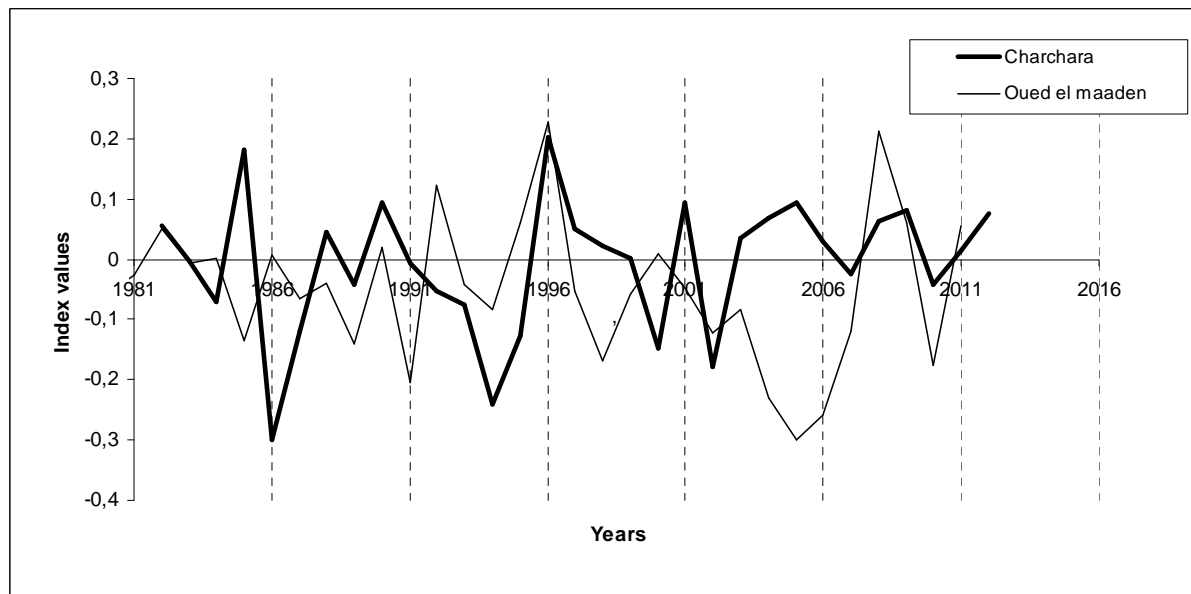
### Chronologies

The trees age averages are 29.6 years in Oued el maaden and 36 years in Charchara. The general appearance of chronologies averages by sector (Figure. 6) reveals a rather irregular radial growth in the area of Oued el maaden. Charchara sector presented, at an equal cambial age, growth average significantly higher than Oued el maaden.



**FIGURE 6. MASTER CHRONOLOGIES BY SITE OF PINUS HALEPENSIS POPULATION. THICKNESS OF RAW CIRCLES 1/1000 MM.**

Looking at the normalized curves, we noted the disappearance of any age tendency and the geometrical effect, as well as the attenuation of peaks or the depressions of growth (Figure. 7). We observed that trees from Charchara react stronger than from Oued el maaden of 1981 to 2013 to the local conditions.



**FIGURE 7. NORMALIZED CHRONOLOGIES OF PINUS HALEPENSIS.**

#### IV. DISCUSSION

The impact of heavy metal pollution on ecosystems due to anthropogenic activities like smelting or mining activities has been frequently investigated (Adriano, 1986; Chuan *et al.*, 1996; Cambier, 1997; Dijkstra, 1998; Sheppard *et al.*, 2000). The atmospheric pollutants have affected forests and soils during the last century. Particularly, long range atmospheric transport of heavy metals can lead to pollutant deposition even in supposedly pristine areas (De Vries *et al.*, 2002). This study highlights the impact of dust career emissions on the growth and productivity of Aleppo pine forest.

Soil analyses showed a highly clay content in Oued el maaden than charchara site (Figure. 5). In fact, clays, with their physico-chemical properties, play a very important role in the availability of heavy metals. Indeed, heavy metals can be absorbed and immobilized by clay minerals (Li, 2000). The clay proportion has a great importance for metal behaviour in the different soils and plays a major role in the physico-chemical processes and to metal retention in soils (Hernandez *et al.*, 2003).

The average contents of organic matter O.M under *Pinus halepensis* was  $\pm$  the same in both sites (Table. 1). This result refutes correlations of this parameter with the average tree height and diameter growth. The relationships between organic matter content and heavy metals found in our soil samples are weak. Harter (1983) and McBride *et al.* (1997) suggest that the absence of correlation between total soil organic matter content and heavy metal absorption can be attributed to the fact that the reactive fraction of the organic matter could not be assessed in the relationship. It is possible that decreased tree growth resulted from Pb-induced changes in the soil (Selonen *et al.*, 2014) and concomitant changes in soil processes. It is also possible that Pb impaired tree growth by damaging tree roots and root-associating mycorrhizal fungi (Malkonen *et al.*, 1999; Hartley-Whitaker *et al.*, 2000; Kukkola *et al.*, 2000; Menon *et al.*, 2007; Sousa *et al.*, 2014).

For evaluation of environmental pollution potential effects, content of metal trace ETM were selected as a basic indicator of anthropogenic influence. Table 2 shows the distribution of Pb, Zn and Cd concentrations in the forest soils of the two Aleppo pine populations studied. It is evident that there is a general regularity in this respect: minimum concentrations are occurring in Charchara soil, while the maximum concentrations are in the career areas of the Oued el maaden region.

Human activities contribute largely to the pollution of Environmental matrices by Pb. This was confirmed in the low Pb traces found in the soil samples from the Charchara station (located at a distance of 15 km to the career). Indeed, this soil has

been contaminated by dusts from the mine (Sirven, 2006). Soil Pb contamination is thus rather the result of many years of deposition enriched in transboundary leaded particles of local or regional industrial sources. The soil surface reflects several years of atmospheric deposition. This study highlights that atmospheric pollution inputs have impacted the soils in forest areas, and that soil surface accumulates heavy metals in areas where pollutant inputs are significant but also in more remote areas. Likewise, for water analyses, traces of pollution were highly significant in Oued el maaden site; despite the low concentrations compared in the soil (Table .3).

Tree radial-growth models are valuable for simulating the impacts of environmental conditions changes on the future growth of forest species (Jorge *et al.*, 2012). Wood anatomical features in tree rings have been interpreted as indicators of environmental change (Briffa *et al.*, 2003). In the present study, dendrochronological techniques have been applied to estimate the effects of career pollution on two Tunisian Aleppo pine populations growing under different conditions. Since both sites are in the same climatic, altitudinal and soil conditions (in the Beja region), the difference in radial growth is attributed to the presence of the career. Charchara sector presented a growth average significantly higher than Oued el maaden (Figure 6 and 7). This dendrochronological investigation has demonstrated that the growth dynamics of Aleppo pine forests are affected by a complexity of environmental factors, which mostly involve atmospheric pollutants (Battipaglia *et al.*, 2007). Our study shows that not only the climatic factors are critically important for the growth of trees, but also, Human activities such as mining.

## V. CONCLUSION

The reaction of forests to activities such as human activities gains more and more importance under future. We found considerable differences in overall growth yield of the two sites. Furthermore our findings suggest that the response of trees depends on the local conditions. We showed that atmospheric pollutants induced changes in the soil, water resources and growth of pine trees. In addition, our findings suggest that some crucial functions such as decomposition of organic matter and primary production strongly depend on the contamination history of the site. The Aleppo pine growth is sensitive to environmental changes and can provide clear evidences of pollution impacts on this ecosystem.

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## REFERENCES

- [1] Adriano DC., 1986. Trace elements in the terrestrial environment. New York. Springer, p. 876.
- [2] Battipaglia G., Cherubini P., Saurer M., Siegwolf R.T.W., Strumia S., Cotrufo M.F., 2007. Volcanic explosive eruptions decrease tree ring growth but not photosynthetic rates in the surrounding forest. *Global Change Biol*, 13 : 1122–1137.
- [3] Biondi F., 1993. Climatic signals in tree rings of *Fagus sylvatica* L. from the central Apennines, Italy. *Acta Oecologica*, 14: 57-71.
- [4] Briffa K.R., Osborn T.J., Schweingruber F.H., 2003. Large-scale temperature inferences from tree rings: a review. *Global and Planetary Change*, 40: 11-26.
- [5] Cambier P., 1997. Evaluation of the mobility of toxic elements in contaminated soils. *Anal Mag*, 25: 35-38.
- [6] Chuan M.C., Shu G.Y., Liu J.C., 1996. Solubility of heavy metals in a contaminated soil: effects of redox potential and pH. *Water Air Soil Pollut*, 90: 543-556.
- [7] Corwin DL., Davis A., Goldberg S., 1999. Mobility of arsenic in soil from the Rocky Mountain Arsenal area. *J Contam Hydrol*, 39: 35-58.
- [8] Cotrufo M.F., Virzo de Santo A., Alfani A., Batoli G., De Crisofaro A., 1995. Effects of urban heavy metal pollution on organic matter decomposition in *Quercus ilex* L. *Environ Pollut*, 89: 81-7.
- [9] Couhert B., Duplat P., 1993. Le Pin d'Alep dans la région Provence-Alpes-Côte d'Azur proposition pour une sylviculture et un modèle de production. *Bulletin Technique de l'ONF*, 25 : 3-22.
- [10] Dagnelie P., 1986. Towards greater internationalization of the Biometric Society. *Biom. Bull.* 3 (1), 1-3.
- [11] De Vries W., Vel E., Reinds G.J., Deelstra H., Klap J.M., Leeters E.E.J.M., Hendriks C.M.A., Kerkvoorden M., Landmann G., Herkendell J., Haussmann T., Erisman J.W., 2002. Intensive monitoring of forest ecosystems in Europe 1. Objectives, set-up and evaluation strategy. *Forest Ecol Manag*, 589: 1-19.
- [12] Direction Générale des Forêts., 1995. Résultats du premier inventaire forestier national en Tunisie.
- [13] Dijkstra E., 1998. A micromorphological study on the development of humus profiles in heavy metal polluted and non-polluted forest soils under Scots pine. *Geoderma*, 82: 341-358.



- [14] Friedland A.J., Johnson A.H., 1985. Lead distribution and fluxes in a high-elevation forest in northern Vermont. *J Environ Qual*, 14: 332-6.
- [15] Fritts H.C., 1976. Characteristics of tree rings as predictors of climate. *Abstracts of Papers of the American Chemical Society*, 172, 30.
- [16] Harter R.D., 1983. Effect of soil Ph on adsorption of lead, copper, zinc, and nickel. *Soil Sci Soc Am J*, 47: 47-51.
- [17] Hartley-Whitaker J., Caimey J., Meharg a., 2000. Toxix effects of cadmium and zinc on ectomycorrhizal colonization of Scots pine (*Pinus sylvestris* L.) from soil inoculum. *Environ. Toxicol. Chem*, 19 : 694-699.
- [18] Hawkins J.L., Sheppard M.I., Jorgensen S.S., 1995. Predicing soil lead migration: how can ancient church roofs help?. *Sci Total Environ*, 166: 43-53.
- [19] Hernandez L., 2003. Dynamique des éléments traces métalliques dans les sols de différents écosystèmes forestiers Français : origine, distribution physique et chimique et facteurs de contrôle. Thèse, Université Paul Sabatier, Toulouse, in press.
- [20] Hou H., Takamatsu T., Koshikawa M.K., Hosomi M., 2005a. Migration of silver, indium, tin, antimony, and bismuth and variations in their chemical fractions on addition to uncontaminated soils. *Soil Sci*, 170: 624-93.
- [21] Huang X., Olmez I., Aras K., 1994. Emission of trace elements from motor vehicles: potential marker elements and source composition profile. *Atmos Environ*, 28: 1385-91.
- [22] Jorge O., Bogino S., Spiecker H., Bravo F., 2012. Climate impact on growth dynamic and intra-annual density fluctuations in Aleppo pine (*Pinus halepensis*) trees of different crown classes. *Dendrochronologia*, 30: 35-47.
- [23] Kukkola E., Rautio P., Huttunen S., 2000. Stress indications in copper- and nickel- exposed Scots pine seedlings. *Environ. Exp. Bot*, 43: 197-210.
- [24] Li W., Gao J.X., 2002. Acid deposition and integrated zoning control in China. *Environmental Management*, 30(2): 169–182.
- [25] Malkonen E., Derome J., Fritze H., Helmisaari H., Kukkola M., 1999. Compensatory fertilization of Scots pine stands polluted by heavy metals. *Nutr. Cycl. Agroecosyst*, 55: 239-268.
- [26] McBride M., Sauvé S., Hendershot W., 1997. Solubility control of Cu, Zn, Cd and Pb in contaminated soils. *Eur J Soil Sci*, 48: 337-346.
- [27] Menon M., Hermle S., Guenthardt-Goerg M.S., Schulin R., 2007. Effects of heavy metal soil pollution and acid rain on growth and water use efficiency of a young model forest ecosystem. *Plant Soil*, 297: 171-183.
- [28] Mizohata A., Ito N., Kusuya Y., 2000. Characterization of the atmospheric particulate matter observed at the sites near the heavy traffic road in Tokyo. *J Jpn Soc Atmos Environ*, 35: 77-102. (in Japanese).
- [29] Munaut A.V., 1978. Première étude palynologique du gisement paléolithique de Biache-Saint-Vaast (Pas- de-Calais). *Bull. Ass. Fr. Et. Quat*, 18: 183-192.
- [30] Nriagu J.O., Pacyna J.M., 1988. Quantitative assessment of worldwide contamination of air, water and soils by trace metals. *Nature*, 333: 134-9.
- [31] Nriagu J.O., 1990a. Global metal pollution. *Environment*, 32: 28-33.
- [32] Padilla K.L., Anderson K.A., 2002. Trace element concentration in tree rings biomonitoring centuries of environmental change. *Chemosphere*, 49: 575-585.
- [33] Ponette Q., Ulrich E., Brêthes A., Bonneau M., Lanier M., R., 1997. RENECOFOR : Chimie des sols dans les 102 peuplements du réseau. Office National des Forêts, Département des Recherches Techniques, p 427.
- [34] Poszwa A., Ferry B., Dambrine E., Wickman T., Loubt M., Bishop K., 2004. Variations of bioavailable Sr concentration and Sr87 / Sr86 ratio in boreal forest ecosystems – role of biocycling mineral weathering and depth of root uptake. *Biogeochemistry*, 67: 1-20.
- [35] Quezel P., 1986. Les pinus du groupe 'halepensis'. *Ecologie Végétaion, Ecophysiologie, Options méd-uterranéennes*. C.I.H.E.A.M. Vol.1.
- [36] Royal Ministry for Foreign Affairs; Royal Ministry of Agriculture (Eds.), 1971. Air Pollution Across National Boundaries. The impact of sulphur in Air and Precipitation, Sweden's Case Study for the United Nations Copnference on the Human Environment. Stockholm.
- [37] Ruan X.I., Zhang G.I., Ni L.J., He Y., 2008. Distribution and migration of heavy metals in undisturbed forest soils : a high resoltion sampling method. *Pedosphere*, 18: 386-93.
- [38] Sakata M., Marumoto K., Narukawa M., Asakura K., 2006. Regional variations in wet and dry deposition fluxes of trace elements in Japan. *Atmos Environ*, 40: 521-31.
- [39] Selonen S., Liiri M., Setälä H., 2014. Can the soil fauna of boreal forests recover from lead-derived stress in a shooting range area? *Ecotoxicology*, 23: 437-448.
- [40] Sheppard D.S. Claridge G.G.C., Campbell I.B., 2000. Metal contamination of soils at Scott base, Antartica. *Appl Geochem*, 15: 513-530.
- [41] Shparyk Y.S., Parpan V.I., 2004. Heavy metal pollution and forest health in the Ukrainian Carpathians. *Environmental Pollution*, 130: 55-63.
- [42] Sirven J.B., 2006. Détection de métaux lourd dans les sols par spectroscopie d'émission sur Plasma induit par laser (LIBS). Université Bordeaux I. Thèse. 253p.
- [43] Smith W.H., 1981. Air Pollution and Forests. Interaction between Air Contaminants and Forest Ecosystems. New York, Heidelberg, Berlin.

- [44] Sousa N.R., Ramos M.A., Marques A.P.G.C., Castro P.M.I., 2014. A genotype dependent-response to cadmium contamination in soil is displayed by *Pinus pinaster* in symbiosis with different mycorrhizal fungi. *App. Soil Ecol*, 76: 7-13.
- [45] Takamatsu T., Takada J., Matsushita R., Sase H., 2000. Aerosol elements on tree leaves – antimony as a possible indicator of air pollution. *Glob Environ Res*, 4: 49-60.
- [46] UN/ECE., 1979. Convention on Long-range Transboundary Air Pollution. Geneva.
- [47] Vila B., Vennetier M., Ripert C., Chandioux O., Liang E., Guibal F., Torre F., 2008. Has global change induced opposite trends in the radial growth of *Pinus sylvestris* and *Pinus halepensis* at their bioclimatic limit? The example of the Sainte-Baume forest (South-east France). *Annals of Forest Science*, 65 (709).
- [48] Wang Y.F., Huang K.L., Li C.T., Mi H.H., Luo J.H., Tsai P.J., 2003. Emissions of fuel metals content from a diesel vehicle engine. *Atmos Environ*, 37: 4637-43.